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Key Points:
- Prescribed fires changed 2020 wildfire burn severity by −16% in the western US and smoke emissions by −101 kg per acre in California
- Fire treatments in the wildland-urban interface were less effective at reducing burn severity and smoke emissions than those outside it
- Overall, prescribed fire use led to a net reduction of −14% in smoke emissions, considering contributions from both wildfires and treatments

Supporting Information:
Supporting Information may be found in the online version of this article.

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Effect of Recent Prescribed Burning and Land Management on Wildfire Burn Severity and Smoke Emissions in the Western United States

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Abstract Wildfires in the western US increasingly threaten infrastructure, air quality, and public health. Prescribed ("Rx") fire is often proposed to mitigate future wildfires, but treatments remain limited, and few studies quantify their effectiveness on recent major wildfires. We investigate the effects of Rx fire treatments on subsequent burn severity across western US ecoregions and particulate matter ($PM_{2.5}$) emissions in California. Using high-resolution (30-m) satellite imagery, land management records, and fire emissions data, we employ a quasi-experimental design to compare Rx fire-treated areas with adjacent untreated areas to estimate the impacts of recent Rx fires (Fall 2018–Spring 2020) on the extreme 2020 wildfire season. We find that within 2020 wildfire burn areas where Rx fires were used prior to 2020, burn severity changed by −16% ($p < 0.001$) and smoke $PM_{2.5}$ emissions changed by −101 kg per acre ($p < 0.1$). Rx fires in the wildland-urban interface ("WUI") were less effective in reducing burn severity and smoke $PM_{2.5}$ emissions than those outside the WUI. Overall, Rx fires led to a net reduction of −14% in $PM_{2.5}$ emissions, including those from the Rx fires themselves. The proposed policy of treating one million acres annually in California could reduce smoke emissions by 655,000 tons over the next 5 years, equivalent to 52% of the emissions from 2020 wildfires. Our analysis provides comprehensive estimates of the net benefits of Rx fire on subsequent burn severity and smoke $PM_{2.5}$ emissions in the western US, an empirical basis for evaluating proposed Rx fire expansions, and valuable constraints for future modeling.

Plain Language Summary Prescribed ("Rx") fire is increasingly proposed as a policy strategy to reduce wildfire risks in the western US, but evidence of its effectiveness in lowering fire severity and smoke emissions remains limited. We empirically demonstrate that, for areas that had been recently treated with Rx fire and then burned during the extreme 2020 wildfire season in California, Rx fires led to a net reduction of −14% in smoke emissions, although these treatments were less effective in areas near human populations. Our findings suggest that expanding Rx fire use can meaningfully reduce smoke emissions, even when factoring in smoke from the Rx fires themselves. The broader application of Rx fires can provide benefits in mitigating severe wildfire impacts and improving air quality in similar fire-prone regions worldwide.

1. Introduction

Due to a warming climate, a legacy of fire suppression, and population growth in the wildland-urban interface ("WUI"), the western United States has seen a recent rise in extreme wildfire seasons (Abatzoglou et al., 2021; Anderegg et al., 2022; Rao et al., 2022). Large wildfires can irreversibly alter ecosystems (Stevens-Rumann et al., 2018), destroy human-built environments (St. Denis et al., 2023), and cause poor air quality and health problems due to smoke particulate matter ($PM_{2.5}$) (Burke et al., 2023; Zhou et al., 2021). Prescribed ("Rx") burning is increasingly proposed as a mitigation strategy to reduce the risk and intensity of future wildfires. In the US, this includes a national investment of nearly $2 billion toward the reduction of hazardous fuels using Rx burns and other treatments (H.R.5376, 2022) and a California plan to treat 400,000 acres by the end of 2025 with a broader objective to treat one million acres annually across the state (California's Wildfire and Forest Resilience Action Plan, 2021). However, there is limited systematic, quantitative evidence of the efficacy of Rx burning in reducing fire severity and overall smoke $PM_{2.5}$ emissions.

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Despite the potential benefits of Rx fires to reduce future wildfire severity and smoke, their implementation in the western US remains limited (Kelp et al., 2023). While Indigenous fire practices and strategies across the US demonstrate the advantages of Rx fires for ecosystem management (Adlam et al., 2022; El Asmar et al., 2024; Lake & Christianson, 2019), public acceptance in the western US is hindered by concerns over smoke impacts and escaped fires (Kolden, 2019). Additionally, climate warming has reduced the burn windows for safe Rx burning, complicating efforts to manage wildland fire risks (Jonko et al., 2024; Swain et al., 2023). The primary policy focus of Rx fire management in the western US has been to protect communities in the WUI, which presents issues ranging from efficacy to equity (Keiter, 2012). The spread of homes into wildfire-prone areas (United States Government Accountability Office, 2019) amplifies these risks. Wealth disparities mean that while wealthier homeowners may afford home-hardening measures, poorer districts struggle with the associated costs (Auer, 2021). Furthermore, although Rx fires generally produce less smoke and have higher combustion efficiency on average compared to wildfires (Marsavin et al., 2023; Schollaert et al., 2024), Rx fires can still negatively impact air quality and disproportionately affect vulnerable communities (Afrin & Garcia-Menendez, 2021). In contrast to mechanical thinning, which primarily reduces canopy density and removes smaller ladder fuels that contribute to crown fire behavior, Rx fire consumes litter and understory shrubs, thereby reducing future fire intensity (Brodie et al., 2024). Existing research lacks a clear method to quantify the trade-offs between Rx fires and future wildfire risk reduction, leaving a gap in understanding the overall benefits versus the potential public health costs.

Evidence on the net effects of Rx burning in the western US is limited and primarily derived from a small number of case studies conducted before the 2018 wildfire season. Globally, most studies on Rx fires take place in North America (Fuhlendorf et al., 2011), with additional studies focusing on regions in Australia (Boer et al., 2009; Collins et al., 2023), the Mediterranean (Fernández-Guisuraga & Fernandes, 2024), and Africa (Sow et al., 2013). These works include characterizing Rx fire effects on wildfire smoke emissions (Hunter & Robles, 2020) and severity (Davis et al., 2024), but none assess empirically the impact of Rx fires on burn severity and smoke emissions from subsequent wildfires. A recent meta-analysis by Davis et al. (2024) examined 40 publications evaluating wildfire severity in both Rx fire-treated areas and untreated controls for wildfires spanning from 1994 to 2016 and the Dixie Fire in 2021 in the western US (Davis et al., 2024). Using mixed severity metrics (e.g., crown scorch height, percent canopy cover change, burn severity derived from satellite imagery), they found that Rx burns reduced severity by 62% relative to untreated areas. Most of these experimental designs compared fire risks, severity, and intensity between areas treated with Rx fires and untreated areas, accounting for variations in fire weather, slope, topography, and land cover types. However, these studies did not include information about smoke emissions, treatment sizes, and other environmental covariates such as proximity to the WUI. The WUI, where undeveloped wildland vegetation and human development meet, is the area where fires pose the greatest risk to people due to the proximity of flammable vegetation (Radeloff et al., 2018). Most observational studies occur at the scale (~1,000 m^2) of a forest canopy (e.g., Vaillant et al., 2009), with few results addressing $PM_{2.5}$ smoke during recent severe wildfire seasons.

Data on Rx fires are limited, so a variety of assumptions are made that lead to potential spatiotemporal discrepancies. Low-intensity fires are often used as proxies for Rx fire treatments (Wu et al., 2023), although these fires are frequently ignited by lightning as opposed to humans (Rao et al., 2023) and generally have different seasonal trends (Coogan et al., 2020). Lightning-ignited fires tend to occur more frequently with convective events such as thunderstorms and over specific orographic features such as mountain ranges (Soler et al., 2021) but are relatively random with respect to proximity to the WUI (Dorph et al., 2022). In contrast, Rx fire planning typically has specific at-risk communities in mind (Keiter, 2012). In modeling Rx fires, few observational constraints exist, requiring studies to rely on historical projections of Rx fires (Kramer et al., 2023) or to create hypothetical case studies (Kiely et al., 2024; Rosenberg et al., 2024). Moreover, most regional modeling efforts use resolutions greater than 10 km even though most Rx fires cover less than 100 acres (approximately 0.4 km^2), underscoring the need for high-resolution analysis.

We empirically assess the effects of Rx fire treatments on burn severity in the western US and $PM_{2.5}$ emissions in California during the extreme 2020 wildfire season. We use 30-m high-resolution satellite imagery, historical land management records, and detailed wildfire and Rx fire emissions inventory data. We develop a quasi-experimental design to compare Rx fire-treated areas with adjacent control areas defined in this study. We define treated areas based on Rx fire records from Fall 2018 to Spring 2020 in areas that subsequently burned in wildfires in 2020. To compare outcomes, we designate nearby untreated "control areas" outside the treated zones

Table 1
Characteristics of 2020 Wildfires Overlapping With NFPORS Treatments

Wildfire	State	Ignition date	Acres burned	NFPORS acres treated (number of treatments)
August Complex	CA	2020-08-17	1,032,648	1,716 (73)
Bobcat	CA	2020-09-06	115,997	217 (5)
Bush	AZ	2020-06-13	193,455	534 (1)
Cameron Peak	CO	2020-08-13	208,663	738 (20)
Creek	CA	2020-09-05	379,895	1,519 (81)
East Fork	UT	2020-08-21	89,568	1,909 (10)
Lake	CA	2020-08-12	31,089	8 (1)
Mangum	AZ	2020-06-08	71,450	7,814 (6)
Medio	NM	2020-08-17	3,775	43 (1)
Mullen	WY	2020-09-17	176,878	342 (5)
Phillips Creek	ID	2020-08-05	2,112	552 (1)
Sheep	CA	2020-08-17	29,570	668 (7)
Slater	CA/OR	2020-09-08	157,220	872 (41)
Superstition	AZ	2020-08-20	9,539	183 (3)

of equal size. We then quantify whether subsequent burn severity and smoke $PM_{2.5}$ emissions during 2020 wildfires differed between treated and control areas, using a regression approach that controls flexibly for land cover type, past fire activity, and whether sampled pixels were in the WUI (Materials and Methods). In essence, our approach assumes that absent treatment, a pixel treated with Rx fire would have had the same burn severity and $PM_{2.5}$ emissions as a nearby untreated pixel, conditional on the controls. Finally, we estimate the net effect on $PM_{2.5}$ emissions per acre burned by Rx fires in California—that is, the tradeoff between additional emissions from Rx fire and reduced emissions from subsequent wildfires—along with the implications for the type of dramatic near-term scaling of Rx fire efforts that is currently being proposed in the state.

2. Materials and Methods

2.1. Rx Fire and Land Management Data Sets

The National Fire Plan Operations and Reporting System (NFPORS) fuels treatment database is maintained by the US Department of the Interior (DOI) collaboratively with the US Department of Agriculture (DOA). NFPORS reports Rx fires with a resolution as fine as 1 acre (~0.004 km^2). It records whether a treatment is accomplished in the WUI, the size of the treatment in acres, the category of treatment (e.g., Rx fire, mechanical thinning), along with unique treatment IDs. Our analysis is focused on the 2020 extreme wildfire season. We use historical records of Rx burn locations from October 2018, when comprehensive geolocated data on Rx burned areas first became available, through May 2020, using Rx burns that overlap with subsequent wildfires during the 2020 wildfire season (July–November). Starting in 2018, these data are available as point data and an accompanying acreage (but do not contain treatment polygons). For wildfires, we use the Monitoring Trends in Burn Severity (MTBS) (Eidenshink et al., 2007) database, which uses 30-m Landsat imagery to define the final fire (polygon) perimeters and assess burn severity for all fires over 1,000 acres (~4 km^2) in the western US (Picotte et al., 2020). We find that 255 NFPORS treatments intersect with 14 wildfires, with 6 of these wildfires located in California (Table 1). After removing overlapping treatment locations in space, we have 186 unique NFPORS treatments (average size of 55 acres).

NFPORS did not extensively report geolocated information on burned areas before the fall of 2018 and only provided longitude and latitude information without final treatment perimeters. As a result, we construct random sampling strategies (detailed in a following section) to estimate the effects of land management treatment in the absence of provided perimeter information. Additionally, we compare our data to the Rx fire perimeters data set (https://map.dfg.ca.gov/metadata/ds0397.html) from California Department of Forestry and Fire Protection (CAL FIRE). The CAL FIRE data set includes perimeters from multiple agencies and provides associated data such as project number, start date, and acres reported. However, the CAL FIRE data set reports a fraction of the treatments conducted by the DOI and DOA. For example, NFPORS reports 115 unique treatments within the Creek Fire perimeter between 2018 and May 2020, while CAL FIRE reports only 36 treatment perimeters, despite all treatments being conducted by or in collaboration with the DOI and DOA. The NFPORS data set reports general treatment types (e.g., Fire ($n = 115$), Mechanical ($n = 60$), and "Other" which is largely chemical treatments ($n = 11$)) as well as subtypes for specific land management techniques: machine pile burn, broadcast burn, biomass removal, thinning, crushing, fire use, lop-and-scatter, and chemical treatments (samples sizes found in Table S3 in Supporting Information S1). While these treatment subtypes are important for understanding which techniques result in more effective reductions in fire severity and smoke $PM_{2.5}$ emissions, we focus on general treatment types due to greater statistical power and balanced sample sizes. Nevertheless, we provide coefficient estimates for these specific techniques, divided into areas inside and outside the WUI, in Table S3 in Supporting Information S1. The Rx fire treatments we report here may include mixed methods, such as mechanical thinning followed by burning (e.g., pile burning), whereas the mechanical treatments exclusively omit the use of fire.

2.2. Satellite Data Sets

We employ a burn severity gridded data set derived from the Sentinel-2A satellite from the European Space Agency. We use the Google Earth Engine (GEE) cloud computing platform (Gorelick et al., 2017), which hosts Sentinel-2 Level 2A data containing 13 spectral bands with spatial resolutions ranging from 10 to 60 m. We retrieve imagery from 2 weeks before and 2 weeks after a wildfire occurrence, as determined by MTBS perimeters and ignition dates. We exclude pixels with a greater than 65% probability of being obscured by cloud cover using the Sentinel-2 cloud probability 10 m data set on GEE. For each pre- and post-fire image, we calculate the Normalized Burn Ratio (NBR), a common spectral index for fire severity that approximates the burn effects by dividing the difference between the near-infrared (NIR; 835 nm) and shortwave infrared (SWIR; 2022 nm) central bands by their sum (Miller & Thode, 2007). We then calculate the differenced Normalized Burn Ratio (dNBR), which quantifies the fire-induced changes in vegetation greenness and landscape moisture content, by subtracting the post-fire NBR from the pre-fire NBR:

$$\mathrm{dNBR} = \left(\frac{\mathrm{NIR}_{\mathrm{pre-fire}} - \mathrm{SWIR}_{\mathrm{pre-fire}}}{\mathrm{NIR}_{\mathrm{pre-fire}} + \mathrm{SWIR}_{\mathrm{pre-fire}}}\right) - \left(\frac{\mathrm{NIR}_{\mathrm{post-fire}} - \mathrm{SWIR}_{\mathrm{post-fire}}}{\mathrm{NIR}_{\mathrm{post-fire}} + \mathrm{SWIR}_{\mathrm{post-fire}}}\right) \tag{1}$$

The final data set resolution is resampled to 30 m to match the resolution of the other data sets used in this work. A negative dNBR value or value of 0 indicates no fire effect on vegetation, while increasingly positive dNBR values suggest higher burn severity. All dNBR values less than 0 were excluded from this analysis.

For land cover classifications, we use the 2019 National Land Cover Database (NLCD), which is a Landsat-based data set that uses digital change detection methods to identify changes in land cover, impervious cover, and forest canopy cover across the US (Jin et al., 2023). The data resolution is at 30 m for the year 2019, and we focus on three broad land cover types: forest, shrub, and barren.

For elevation data, we use the NASA Digital Elevation Model (NASADEM), which is also at 30-m resolution (Crippen et al., 2016) and is a reprocessed version of Shuttle Radar Topography Mission data from 2000, with improved height accuracy and filled missing elevation data. Both NLCD and NASADEM data were retrieved and processed in GEE using MTBS perimeters.

2.3. Fire Emissions Data Sets

To estimate $PM_{2.5}$ emissions from wildfire smoke, we use the Wildfire Burn Severity and Emissions Inventory (WBSE). WBSE is a severity-based emissions inventory that uses Landsat imagery to calculate burn severity through dNBR. The Moderate Resolution Imaging Spectroradiometer (MODIS) and Visible Infrared Imaging Radiometer Suite (VIIRS) active fire detections, with spatial resolutions of 1 km and 375 m respectively, are used to determine the day of burning for each pixel. Vegetation types and emission factors are informed by California-specific field studies to calculate smoke $PM_{2.5}$ emissions. WBSE provides a 30-m resolution for event-based emissions in California, covering the six California fires listed in Table 1. Although WBSE is limited to

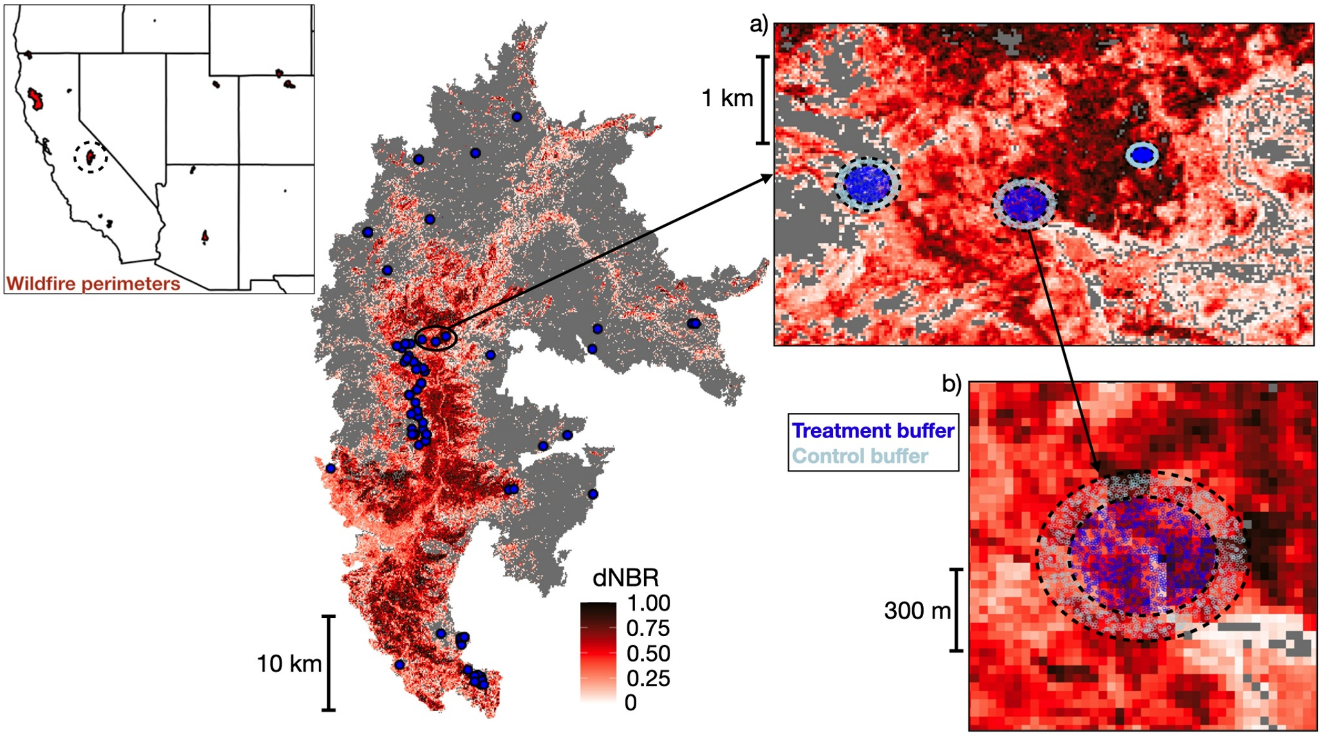

Figure 1. Approach to estimating the impact of Rx fire on burn severity, using the Creek Fire as an example. The Creek Fire perimeter contains 30 m pixels of dNBR values from Sentinel-2 with higher values in dark red indicating more severe burns. Blue dots represent Rx fire treatment locations recorded by NFPORS ($n = 59$) from October 2018 to May 2020. Insets (a, b) show zoomed-in views of our randomly generated treatment buffers centered on the NFPORS coordinates (blue dots) and the surrounding control buffers (cyan dots).

California, it offers the highest resolution smoke $PM_{2.5}$ emissions data with a strong correlation to burn severity metrics.

To estimate $PM_{2.5}$ emissions from Rx fire smoke, we use a reclassified FINNv2.2 source-specific inventory of daily $PM_{2.5}$ emissions from Rx fire across California (Schollaert et al., 2024). Schollaert et al. reclassified the FINN emissions inventory (Wiedinmyer et al., 2023) data by spatially matching it with fire-type information from national and state-level fire and fuel treatment databases, including from CAL FIRE. The Rx fire emissions are provided at a daily 1 km resolution and have been validated using county-level estimates from the EPA's National Emissions Inventory.

2.4. Quasi-Experimental Design Sampling Strategy

To evaluate the effects of Rx fire treatments on burn severity and $PM_{2.5}$ emissions during the 2020 wildfire season, we employ a quasi-experimental sampling design using location data from NFPORS (Figure 1). Our analysis aims to estimate these Rx fire impacts conditional on a wildfire occurring. We identify overlaps between land management areas treated in NFPORS from October 2018 to May 2020 and MTBS wildfire perimeters during the 2020 wildfire season. Based on these intersections, we develop a random sampling strategy to create treatment and control buffers around each set of coordinates, each buffer corresponding to the total acreage treated.

We define the treatment area as a circular buffer centered on an NFPORS coordinate. We then define the control area as a concentric circle completely enclosing the treatment buffer, with its area equal to the treatment acreage but excluding the enclosed treatment buffer area. This design ensures that the control buffer captures areas directly outside the treatment zone while maintaining an equivalent acreage.

We generate 1,000 random points within both the treatment and control buffers to capture the impact inside and outside each Rx fire treatment (Text S1 in Supporting Information S1). For each random point, we extract dNBR values from Sentinel-2A data, $PM_{2.5}$ emissions from WBSE, and covariate information (land cover, elevation).

The random points are seeded to ensure that the burn severity and smoke $PM_{2.5}$ emission impacts at each sampling location are consistent. If there are multiple NFPORS treatments in the same location over time, we report the statistics of the largest treatment in terms of acreage. We acknowledge that Rx fire treatments can occur multiple times in the same location as part of a long-term land management strategy. However, we sample from the largest treatment to avoid double-counting spatially overlapping treatments, ensure a consistent spatial unit of analysis across all sites, and reduce ambiguity in cases where treatments overlapped or were conducted in close succession. No overlapping treatment locations in the data set experienced both Rx fire and mechanical thinning as separate, distinct treatments over time at the same site (Table S4 in Supporting Information S1).

To test for robustness, we increase the size of the treatment and control buffers. We recognize that the control area might still be indirectly affected by the treatment, particularly if the treated area impacts nearby vegetation or other environmental variables. To account for potential spillover effects, we expand the area of both the treatment and control buffers by one-third. Such an adjustment can help to ensure that any treatment effects are distinguished from changes in the control areas. Additionally, to confirm that our method of assigning treated areas by buffering points is reasonable, we use CAL FIRE Rx fire perimeter data to compare treated and control areas within observed Rx fire perimeters.

2.5. Causal Inference of Rx Fire Treatments

We use regression analysis to evaluate the impact of Rx fire treatments on dNBR for all locations and $PM_{2.5}$ emissions for all California fires listed in Table 1. We estimate the following regression:

$$y_{\mathrm{id}} = \beta D_{\mathrm{id}} + \lambda X_{\mathrm{id}} + \alpha_d + \varepsilon_{\mathrm{id}} \quad (2)$$

where y_{id} represents either of our outcomes (dNBR or $PM_{2.5}$ emission) measured at pixel i across our 186 treatment locations d. D_{id} represents a dummy variable for whether a given pixel was treated by an Rx fire treatment, X_{id} is our vector of control variables, which includes indicator variables for whether a given pixel was in the WUI, its land cover type, and whether it had burned in a previous fire, ε is the error term, and α is a vector of dummy variables (separate intercepts, or "fixed effects") for each "treated area" d, which includes both the Rx fire-treated area as well as the surrounding control buffer for a single treatment. The inclusion of treated-area fixed effects ensures that we are only comparing directly adjacent treated and control pixels to one another and not comparing a treated pixel in one location to a distant control pixel. For each regression, we report the 95% confidence interval, where standard errors are clustered at the treatment level. Furthermore, we identify areas treated with Rx fire between October 2018 and May 2020 that previously experienced wildfires between 2001 and 2015. We found five wildfire perimeters (Santiago Fire 2007, Station Fire 2009, Aspen Fire 2013, French Fire 2014, Pickett Fire 2015) that intersected with 38 land management treatments found in NFPORS. For wildfires before 2015, we use Landsat 7 dNBR imagery. Performing similar treatment-control analyses with these buffers indicates that treated areas had a 12.5% increase ($p < 0.001$) in burn severity compared to adjacent controls. To account for past fire history and isolate the effects of Rx fire treatments from legacy impacts, we control for these 38 treatment locations in the above regression.

To test for whether Rx fire treatments have different effects inside or outside the WUI, we first limit our sample to either Rx fire or mechanical thinning treatments and then interact our treatment with an indicator (dummy variable) for whether the treatment was inside the WUI as designated by NFPORS. The coefficient and statistical significance of the estimate on the interaction tell us whether the treatment was larger in the WUI for a given type of treatment; these coefficients are reported in Figure 3c.

To ensure the robustness of our sampling strategy, we perform several additional statistical checks and historical comparisons. We assess the distribution of covariates between treated and control pixels, examining variables such as elevation and land cover types. We conduct t-tests for differences in means and pixel-level regressions to identify significant differences. Covariates showing imbalance between groups are included as controls in the main regression estimates (Figure S3, Tables S5, S6 in Supporting Information S1).

To help ensure that our approach to estimating the impact of Rx fire treatments is actually recovering the impact of treatment rather than random differences in burn severity or emissions that occur within a wildfire burn scar, we implement placebo tests. For each fire, we create 100 random hypothetical treatment locations with accompanying control buffers and compare the distribution of estimated "treatment effects" in these placebo treatment

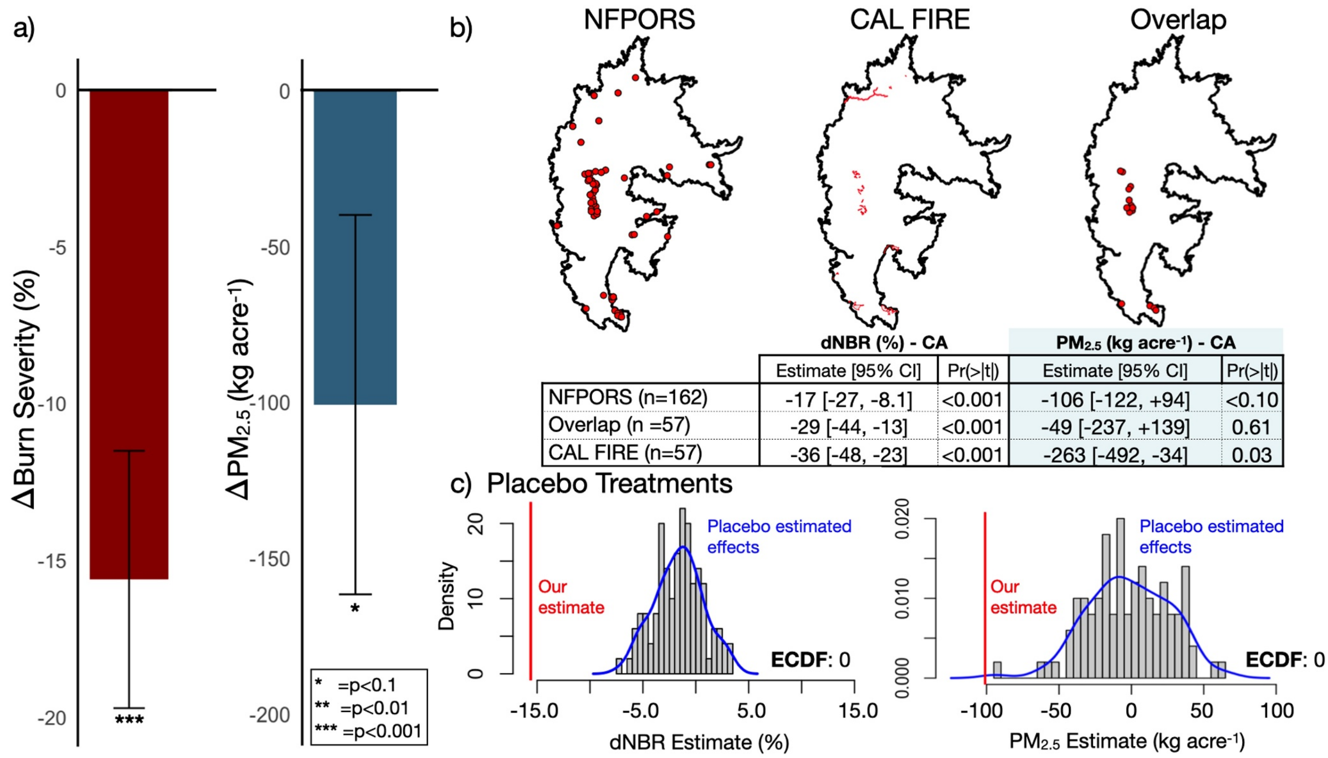

Figure 2. Impact of Rx fire treatments on burn severity in the western US and smoke emissions in California. (a) All sample estimates of burn severity and smoke $PM_{2.5}$ emissions changes in Rx fire-treated areas compared to control areas during the 2020 wildfire season. (b) Comparison of estimates using NFPORS (treatment and control circular buffers), CAL FIRE (treatment perimeters, control circular buffers), and the "overlap" (treatment and control circular buffers) subset of NFPORS inside CAL FIRE perimeters. Maps show overlaps for a single fire (Creek Fire) and the table of estimates shows pooled treatment effect estimates across all fires for which we have data. (c) Results from 100 randomized placebo treatments demonstrate that our estimates of the treatment effect of Rx fires are extremely unlikely to occur by chance ($p < 0.001$). The blue line on the empirical cumulative distribution function (ECDF) plots outlines the distribution density and the red line corresponds to our estimates from (a).

areas to our estimate of the impact of treatment in the true treated area(s) in that same fire. By comparing outcomes ($PM_{2.5}$ and dNBR) of these placebo-treated pixels with actual treated pixels, we can assess whether our observed treatment effects might be attributed to random chance.

To assess the net impact of Rx fire treatments on smoke emissions in California, we use estimates derived from our regression analysis. These estimates allow us to quantify the overall per-acre reduction in smoke $PM_{2.5}$ attributed to Rx fire treatments by accounting for the Rx fire emissions themselves from the reclassified FINNv2.2 inventory from 2012 to 2020. We also identify grid cells where Rx fire emissions occurred in a given year and calculate if they overlapped with any wildfire emission grid cells at a subsequent timestep within a 5 km distance threshold. We then compute the percentage of Rx fire-treated areas that remained unburned. We assume an emissions base year of 2018, which reflects moderate to high wildfire activity. In addition, we compute the total emissions with Rx burning, E_{Rx}:

$$E_{\mathrm{Rx}} = (1-a)\,x + a(x + (1-b)\,y) \tag{3}$$

total emissions without Rx burning, E_{NoRx}:

$$E_{\mathrm{NoRx}} = ay \tag{4}$$

and, the percent reduction in overall smoke emissions by conducting Rx fires:

$$\frac{E_{\mathrm{NoRx}} - E_{\mathrm{Rx}}}{E_{\mathrm{NoRx}}} \tag{5}$$

Here, the x variable is the average emissions from an acre of Rx fire calculated by dividing the total emissions from the FINNv2.2 inventory divided by the acres burned by these fires. The y variable is the average emissions from an acre of wildfire burned, which we calculate by dividing the total emissions from our wildfire case studies (here, the Creek and Slater Fires) by the acres burned in Table 1 for these fires. The a variable is the proportion of Rx fire-treated areas that later reburned described above. The percent reduction in wildfire emissions due to an Rx fire, b, is calculated as follows:

$$b = \frac{y - z}{y} \tag{6}$$

where z is the fire-specific effect of Rx fire treatments on smoke emissions estimates and observed decreases for both the Creek and Slater Fires chosen due to data availability. Because Rx fire treatments in these two fires produced different estimates, we take the weighted average based on acres treated in NFPORS for Creek (1,519 acres) and Slater (872 acres). The $a(1 - b)y$ term describes the overlap of Rx fire and wildfire emissions, accounting for the fact that if an area reburns it will emit a reduced amount of wildfire smoke because Rx fire treatment had already occurred. The $(1 - a)x$ term describes Rx fire emissions in areas that do not later reburn in a wildfire. The ax term corresponds to Rx fire emissions in areas that later experienced wildfire. If Equation 5 is positive, Rx fires result in a net savings of smoke emissions, while a negative value implies that Rx fires contributed more emissions than they mitigated during subsequent wildfires. Finally, we scale up these per-acre emission reductions to align with the target treatment of 1 million acres mandated by California's Wildfire and Forest Resilience Task Force.

3. Results

3.1. Efficacy of Rx Burning in the Western US

When investigating the 2020 wildfire season, we find that Rx fire treatments in the two years prior to a wildfire significantly reduced burn severity in the western US and smoke emissions in California (Figure 2a). On average across the western US, Rx fire-treated areas show a reduction of −16 [−24, −7.6]% ($p < 0.001$) in burn severity compared to control areas. In California, Rx fire treatments lead to a −101 [−220, +18] kg per acre ($p < 0.1$) change in smoke $PM_{2.5}$ emissions, with similar shifts observed in burn severity (−17 [−26, −8.2]%, $p < 0.001$) (Table S1 in Supporting Information S1). Increasing the buffer radius around treatments and controls slightly reduces the magnitude of these estimates but does not alter their direction or statistical significance (Table S2 in Supporting Information S1).

We conduct a number of analyses to test the robustness of these primary results. Figure 2b shows the comparison of our experimental sampling (Figure 1) to more precise Rx fire perimeters from CAL FIRE. Our sampling method creates Rx burn area polygons by generating a circular buffer around the geographic point location based on the reported burn area from NFPORS. This sampling strategy likely mischaracterizes the precise Rx fire-treated area. To understand whether this mis-measurement matters, we use the more precise CAL FIRE perimeters for the more limited set of treatments in those data, constructing adjacent control buffers and estimating treatment effects in the same manner. For this more limited set of perimeters in California, we estimate a reduction in burn severity of −36 [−48, −23]% ($p < 0.001$) and in smoke $PM_{2.5}$ emissions of −263 [−492, −34] kg per acre ($p < 0.1$). If instead of using these precise perimeters we estimated Rx fire treatment effects using our circular buffers at the same locations as CAL FIRE, burn severity is changed by −29 [−44, −13]% ($p < 0.1$) while smoke $PM_{2.5}$ emissions changed by −49 [−237, +139] kg per acre ($p = 0.61$). The discrepancy in $PM_{2.5}$ estimates likely reflects the smoothing effect of emission factors in WBSE, which uses average values by vegetation class and may miss fine-scale variation in fire intensity. While both Rx fire boundaries show statistically significant reductions in burn severity, the CAL FIRE perimeters display a stronger $PM_{2.5}$ effect, likely because burn severity captures finer spatial variation, whereas $PM_{2.5}$ estimates are smoother and more sensitive to boundary precision.

To further understand whether our measured differences in burn severity and $PM_{2.5}$ emissions between treated and adjacent control pixels could have occurred by chance, we run a set of placebo experiments in which, within the same fires, we estimate the "impact" of 100 placebo treatments and compare the distribution of these placebo estimates to our estimate of the true treatment effect of Rx fire (Section 2.5). Figure 2c displays our treatment effect estimate relative to the placebo distribution. For both burn severity and smoke emissions, our treatment

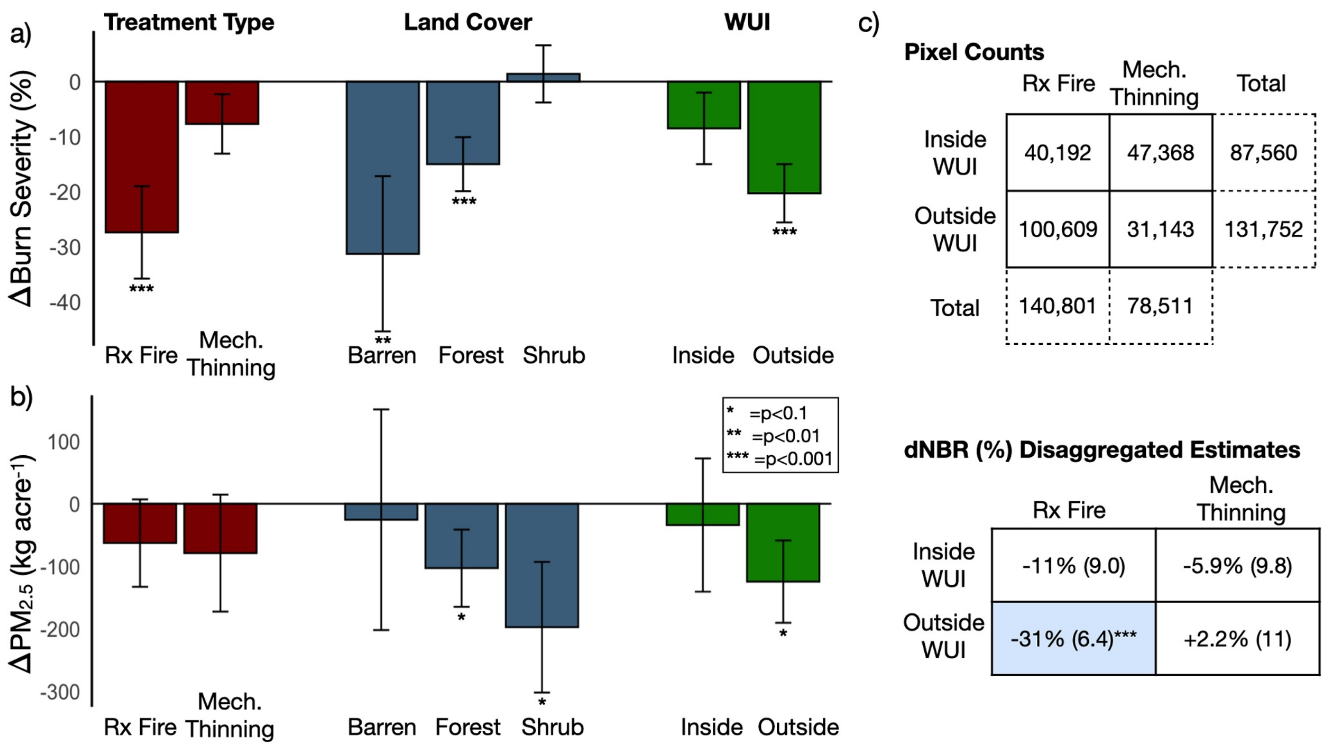

Figure 3. Comparative efficacy of wildfire management strategies. (a) Estimates of burn severity changes in Rx fire-treated buffers compared to control buffers during the 2020 wildfire season in the western US, by treatment type, land cover, and whether the treated area was in the Wildland-Urban Interface ("WUI"). (b) As in (a), but for $PM_{2.5}$ smoke emissions reduction in California. (c) Disaggregated statistics for treatment type (Rx fire vs. mechanical thinning), and inside versus outside of the WUI across the western US.

effect estimate is entirely outside the distribution of placebo treatment effects, which are themselves centered on zero as expected, indicating that our estimated treatment effects are highly unlikely to happen by chance in our data.

3.2. Characterizing Land Treatments in the Western US

Our findings reveal that Rx fire treatments are significantly more effective in reducing burn severity compared to mechanical thinning. Figure 3a shows that across the western US, Rx fire treatments change burn severity by −27 [−44, −10.8]% ($p < 0.001$), whereas mechanical thinning treatments only change burn severity by −7.7 [−18, +2.8]% ($p = 0.15$). These results are consistent with Davis et al. (2024), which found mechanical thinning to be 35% less effective in reducing burn severity in subsequent wildfires than Rx fire treatments. Rx fire consumes a wide range of fuel types, including fine fuels and larger woody debris, whereas mechanical thinning targets larger vegetation and thus often leaves behind smaller fuels (Stephens & Moghaddas, 2005).

In forest ecosystems, land management treatments including Rx fire and mechanical thinning significantly reduce both burn severity in the western US and smoke emissions in California (Figures 3a and 3b). Specifically, these treatments change burn severity by −15 [−25, −5.3]% ($p < 0.001$) and smoke $PM_{2.5}$ emissions by −103 [−224, +18] kg per acre ($p = 0.09$). In barren areas where vegetation accounts for less than 15% of total cover, treatments show a significant reduction in burn severity of −31 [−58, −4.6]% ($p = 0.03$) but the effect on smoke $PM_{2.5}$ emissions is minimal (−26 [−373, +321] kg per acre, $p = 0.89$). In shrublands, the impact of treatments on burn severity is not significant (1.4 [−8.8, +12]%, $p = 0.79$) but there is a significant reduction in smoke $PM_{2.5}$ emissions (−198 [−405, +8.7] kg per acre, $p = 0.06$).

We find that Rx fire treatments are less effective within the WUI compared to outside it (Figures 3a–3c). Treatments inside the WUI change burn severity in the western US by −8.5 [−21, 4.1]% ($p = 0.19$) and change smoke $PM_{2.5}$ emissions in California by −34 [−244, 176] kg per acre ($p = 0.75$). In contrast, treatments outside the WUI show a significant reduction in burn severity of −20 [−31, −10.0]% ($p < 0.001$) and a reduction in smoke $PM_{2.5}$ emissions of −125 [−255, 4.7] kg per acre ($p = 0.06$). On average, the number of acres treated is larger inside than outside the WUI ($p < 0.001$, Figure S1 in Supporting Information S1). Figure 3c indicates that most treatments outside the WUI use Rx fire, while treatments inside the WUI predominantly use mechanical thinning. Statistical tests confirm that Rx fire outside the WUI significantly reduces burn severity, whereas other combinations of WUI designation and treatment type do not.

3.3. Net Rx Burning Effects and Future Projections

We quantify the net impact of Rx fire treatments on smoke emissions, considering both the emissions from Rx fires themselves and subsequent prevented smoke from future wildfires (Section 2.5). Emissions from Rx fires are derived from a reclassified FINNv2.2 source-specific inventory of daily $PM_{2.5}$ emissions, and emissions from wildfires are from the WBSE inventory. We use these data and our results to calculate three quantities: (quantity 1) the ratio of emissions from an average acre of Rx fire versus an average acre of wildfire; (quantity 2) the per-acre reduction in emissions during a wildfire resulting from having conducted a previous Rx fire treatment in an area that subsequently burned, which is used to calculate the emissions benefits of a dramatic near-term scaling of Rx fire efforts that is currently being considered in California (California's Wildfire and Forest Resilience Action Plan, 2021); and (quantity 3) the ratio of total emissions from conducting an Rx burn to total emissions had that burn not happened, accounting for emissions from the Rx burn itself and the probabilistic benefits that burn has on subsequent wildfire emissions. This last ratio is our preferred estimate of the expected net benefits from implementing Rx fire.

We find that the net effects of Rx fires result in overall emission savings, though estimated total savings from observed Rx fires are small given their limited implementation. The Creek and Slater Fires in California contain 66% of all NFPORS treatments in this study and align most closely with observations from the reclassified FINNv2.2 emissions, while other wildfires in California had too few Rx fire observations that overlapped between the data sets. We calculate the fire-specific effect of Rx fire treatments on smoke emissions estimates and observed decreases in both the Creek (−246 kg per acre, $p = 0.07$) and Slater (−293 kg per acre, $p = 0.08$) Fires. Figure 4a shows that the Creek and Slater Fires emitted 213,000 tons of $PM_{2.5}$ smoke. We estimate that the 122 NFPORS treatments occurring prior to these two fires reduced smoke emissions by 630 tons. Inventory estimates suggest the Rx fires at these locations emitted 144 tons of smoke, yielding a net savings of 486 tons of smoke emissions. Although this subset of treatments yields a net smoke savings, the reduced emissions account for only −0.17% of total wildfire emissions (Figure 4a, 4b inset), highlighting that the scale of treatment remains small relative to overall wildfire emissions.

By design, our study considers Rx fires that subsequently burned in a wildfire. Estimating the net emissions effect of future Rx fires, however, requires accounting for the fact that not all Rx-burned locations will subsequently burn in a wildfire, at least in the near term. We calculate that on average 75% of the land treated by Rx fire burns in a wildfire within the next 8 years (Figure 4b, Figure S2 in Supporting Information S1), which agrees well with encounter rates for Northern California (Beidler et al., 2024). We use this value to adjust our estimate of the net emissions savings from Rx fire (Figure 4a). Using this adjustment, we find that Rx fires yield a net savings of 364 tons. Rx fire smoke only constitutes 17% of the smoke emissions from a wildfire in the same areas (quantity 1). We calculate a −34% reduction in wildfire emissions due to an earlier Rx fire (quantity 2, Equation 6). Compared to a counterfactual scenario where no Rx fire treatments are applied (quantity 3, Equation 4), the application of Rx fire (quantity 3, Equation 3) results in a net reduction of −14% in overall $PM_{2.5}$ smoke emissions (quantity 3, Equation 5).

By scaling our net effect of Rx fire treatments per acre, we estimate that treating one million acres of land in California, as mandated by the Governor's Wildfire and Forest Resilience Task Force, would result in 288,000 tons of emissions from the Rx fires themselves. However, over the subsequent 5 years—reflecting a balanced timeframe between our Rx fire burn window (three years; 2018–2020) and our calculation of reburn potential (8 years; 2012–2020)—we estimate that these treatments would reduce emissions in subsequent wildfires by 943,000 tons, resulting in a net reduction of 655,000 tons of $PM_{2.5}$ smoke emissions. We base this projection on a treatment year comparable to 2018, reflecting accumulated fuel loads and moderate to high wildfire activity. These reductions are substantial relative to total emissions in extreme wildfire years like 2020. Figure 4c shows that scaling our net Rx fire effect estimates to one million acres would save more smoke than the emissions from four Creek Fires and two August Complex Fires, the latter of which burned over a million acres. This projected net

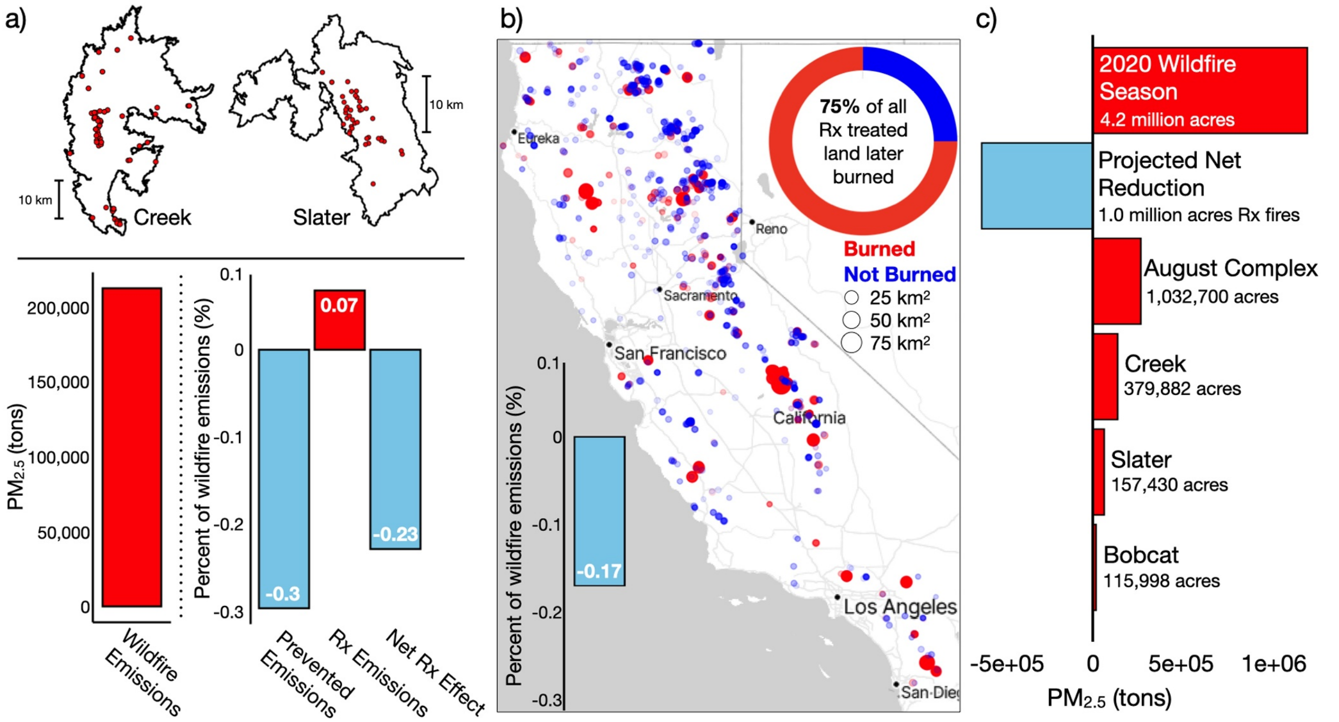

Figure 4. Net effects and projections of Rx fire treatments on smoke emissions in California. (a) The net smoke $PM_{2.5}$ effects from prior Rx fire treatments in the Creek and Slater Fires in terms of both $PM_{2.5}$ emitted from these Rx burns and potential $PM_{2.5}$ saved during these wildfires. (b) The proportion of treated land that subsequently burned in wildfires from a reclassified FINNv2.2 emissions inventory from 2012 to 2020, with an adjusted net smoke $PM_{2.5}$ savings estimate incorporating that, on average, 75% of Rx fire treatments eventually burn. (c) Projecting the potential $PM_{2.5}$ emission reductions if Rx fire treatments are scaled up to one million acres in California by CAL FIRE as mandated by the Governor's Wildfire and Forest Resilience Task Force, with emissions comparisons to other large wildfires during 2020.

reduction includes both the smoke emitted and the smoke saved by Rx fires. The wildfire smoke saved from conducting these Rx fires is the equivalent of 52% of the total emissions from the 2020 wildfire season (conditional on Rx fire-treated areas eventually reburning in a wildfire within a 5-year window).

4. Discussion

Using data on 186 recent Rx fire treatments across the western US, we find that Rx fire treatments effectively reduced burn severity and future smoke emissions from wildfires during the historically active 2020 wildfire season. Our estimates are not driven by differences in land cover or previous fire history between Rx fire-treated areas and adjacent controls, and a placebo exercise indicates that our treatment effects are highly unlikely to arise by chance.

There are at least three reasons why our main estimates could be a lower bound on the benefits of Rx fire on subsequent burn severity and emissions. First, our comparison of NFPORS data and a smaller set of more precise CAL FIRE perimeters (Figure 2b) suggests a more substantial reduction in burn severity and smoke emissions where Rx fire treatments are estimated precisely. However, we cannot rule out the possibility that CAL FIRE treatments differ in some important way from treatments in other locations or jurisdictions. Second, our approach to estimating the treatment effects of Rx fire within subsequently burned wildfire perimeters could underestimate beneficial spillovers from treated areas to neighboring untreated areas, either because treatments reduced severity or emissions in nearby "control" regions that we constructed, or because treatments limited the spatial extent of the wildfire itself. In either case, our approach of comparing treated pixels to neighboring untreated pixels, designed to ensure that these pixels are otherwise similar absent treatment, could lead us to understate the benefits of Rx fire. Finally, to estimate the benefits of substantially scaled Rx fire treatments across California, we account for the fact that not all Rx fire-treated areas subsequently burn in wildfires. However, our calculation of the percentage of Rx fire-treated areas that subsequently burn is based on a limited (8-year) temporal sample and likely underestimates the true probability of near-term reburn. Higher estimates of reburn probability would lead to higher estimated benefits from Rx fire and our calculation of the net reduction in overall smoke emissions are specific to two large, representative wildfires (Creek, Slater) with a sufficient number of reported Rx fire treatments. While our results indicate a net savings in smoke emissions from Rx fires, it should be noted that Rx fires release smoke that can adversely affect human health and disproportionately affect vulnerable communities (e.g., Afrin & Garcia-Menendez, 2021).

Our analysis is limited by the availability of high-resolution emissions data for Rx fires. To address this, we use two complementary smoke emissions data sets: WBSE for wildfire emissions and a reclassified version of FINNv2.2 for Rx fire emissions. These data sets differ in both spatial resolution and methodological approach. WBSE provides 30 m fire-specific emissions estimates calibrated with burn severity and California-specific emission factors (Xu et al., 2022), making it well-suited for our high-resolution, pixel-level fixed effects analysis. In contrast, the reclassified FINNv2.2 is a 1-km emissions product that incorporates fire-type classifications based on federal and state fire and fuel treatment records, allowing us to explicitly distinguish between wildfire and Rx fire in California (Schollaert et al., 2024). We use this data set to estimate the net effects of Rx burning, a calculation that is not currently possible with WBSE.

While these data sets serve complementary roles in our analysis, we acknowledge that they use different fuel consumption assumptions and emission factors. Xu et al. (2022) compared wildfire emissions between WBSE and a modified FINNv1.5 inventory and observed general agreement, although differences could reach up to a factor of three in annual totals. We find a similar level of agreement when comparing smoke $PM_{2.5}$ emissions from WBSE and FINNv2.2 for the California wildfires studied here (Table S7 in Supporting Information S1), with estimates typically falling within a factor of two. Our regression approach controls for key landscape characteristics (e.g., vegetation type, WUI designation), which addresses spatial variation in emissions. However, it does not fully reconcile the underlying methodological differences between the two inventories, which may influence net effect estimates comparing wildfire and Rx fire emissions. $PM_{2.5}$ emission factors (in g kg^{-1}) used in WBSE and FINNv2.2 are generally comparable across temperate forest (10.6 vs. 15.0), shrubland (7.9 vs. 7.1), and grassland (7.2 vs. 7.17) vegetation types. Future work should aim to harmonize the methodologies of wildfire and Rx fire inventories by applying consistent emission factors and fuel assumptions to overlapping fire events.

Moreover, our analysis does not account for the potential effects of vegetation regrowth (dNBR < 0) following Rx fire, which may only be a minor concern over the 2-year post-treatment window analyzed in our causal inference methods but could introduce greater uncertainty when projecting benefits over longer timeframes. Additionally, we do not analyze repeated Rx fire treatments in a single location, opting to sample from the largest treatment to avoid double-counting spatially overlapping areas. Such repeated treatments may be part of long-term land management planning in a region and could either reinforce emissions reductions through sustained fuel removal or diminish effectiveness due to altered fire behavior or fuel composition. Including these overlapping treatments (Table S4 in Supporting Information S1) does not affect the statistical significance of the Rx fire versus mechanical thinning or WUI versus non-WUI comparisons for burn severity (Table S8 in Supporting Information S1). While some marginally significant $PM_{2.5}$ results become insignificant, the WUI relationship remains robust (Table S8 in Supporting Information S1). We note that including these repeated treatments risks overweighting certain areas and assumes equal impact for each intervention, which may not be valid given the smaller scale of these treatments. Isolating the specific effects of each treatment on vegetation in these locations would require high-resolution, time-resolved satellite imagery, which is an important direction for future research. Furthermore, future work incorporating vegetation recovery dynamics and treatment frequency may improve the accuracy of long-term projections and quantify the specific impacts of repeated interventions.

The relatively greater effectiveness of Rx fire in reducing burn severity, compared to mechanical thinning, aligns with previous findings (Davis et al., 2024). This effectiveness is attributed to Rx fire's ability to address a wider range of fuel types and disrupt fuel continuity across landscapes, creating patches of burned and unburned areas that may reduce the spread and intensity of future fires (Figure 3a). In contrast, mechanical thinning primarily targets larger vegetation such as trees and shrubs, often leaving smaller fuels on the ground. While it may reduce vegetation density, mechanical thinning may not create the same level of fuel discontinuity as Rx fire (Agee & Skinner, 2005). We find that land management treatments are more effective in reducing burn severity in forest ecosystems likely due to the heavier fuel loads in forests that typically generate more smoke and heightened burn severity. The effects in barren areas are minimal due to the limited availability of combustible fuel, while shrublands are likely significant in reducing smoke emissions due to the combustion of smaller and more easily ignitable fuels. Further, our study does not account for weather variables at the time of treatment, nor does it differentiate between types of vegetation within land cover categories.

We also acknowledge that the confidence intervals for smoke $PM_{2.5}$ emissions associated with land management treatments include effects greater than zero. This finding suggests that Rx fire treatments may, in some cases, lead to increases in smoke $PM_{2.5}$ emissions, although we find that this possibility is less likely for Rx fire treatments outside the WUI, where our confidence intervals do not contain positive values. More broadly, this finding highlights the inherent tradeoff of Rx fire: while intended to support ecosystem management, Rx fire itself produces smoke. Moreover, due to planning constraints, such as narrow burn windows and regulatory or political considerations, treatment locations may not always be optimized for maximizing smoke $PM_{2.5}$ emissions reductions (Deak et al., 2024).

The reduced effectiveness of Rx fire within versus outside the WUI highlights the challenges of implementing effective Rx fire in areas with dense human populations and infrastructure. There may be several factors related to the WUI that are not fully understood or captured here, which could limit the impact of Rx fires in these areas. These factors might include the application of Rx fire mixed with other methods such as thinning, the weather conditions at the time of ignition, and National Environmental Policy Act (NEPA) mitigation requirements. Moreover, the need to adopt extremely cautious approaches, due to factors concerning community smoke exposure, the risk of escaped Rx fires, and the higher density of structures, could further reduce the treatment's overall effectiveness in the WUI.

The net effects of Rx fire treatments estimated in our analyses indicate potential emission savings, accounting for both smoke $PM_{2.5}$ emissions of Rx fire and prevented smoke $PM_{2.5}$ emissions from future wildfires (Figure 4). While the current scale of Rx fire treatments in the western US is relatively small, California plans to scale up to treating 400,000 acres annually using Rx fire by the end of 2025. This goal, shared among state, federal, tribal, and local entities, is part of a broader objective to treat one million acres annually across California (California's Wildfire and Forest Resilience Action Plan, 2021). Meeting this goal may be challenging, as CAL FIRE treated on average only 30,000 acres annually with Rx fire from 2018 to 2023 (https://www.fire.ca.gov/our-impact/statistics, last accessed: 27 August 2024), which is only 7.5% of its 400,000 acres goal. However, if the goal is met, the smoke savings are likely to be substantial. Not only do our analyses suggest that such a program is likely to reduce a large fraction of the smoke $PM_{2.5}$ emissions in California (Figure 4), but the smoke savings achieved in California may also represent a significant reduction in wildfire smoke exposure across the western US, given the importance of California as a source of wildfire smoke for other regions (Kelp et al., 2023; Wen et al., 2023).

5. Conclusions

We construct a quasi-experimental design that combines 30 m satellite imagery, land management records, and fire emissions data to examine the effects of prior Rx fire treatments during the 2020 wildfire season. We find that, regardless of varying sensitivity definitions, Rx fire treatments conducted within 2 years before a wildfire significantly reduced burn severity and smoke $PM_{2.5}$ emissions. Additionally, land management treatments using Rx fire were significantly more effective at reducing burn severity compared to mechanical thinning in the western US. However, treatments in the WUI predominantly relied on mechanical thinning, which was less effective than Rx fire use. Statistical tests confirm that the limited Rx fires conducted in the WUI did not effectively reduce burn severity, which likely reflects the cautious approaches adopted near populations and infrastructure, despite the WUI being a priority area of policy focus. Furthermore, Rx fires achieved a net reduction of −14% in smoke $PM_{2.5}$ emissions, accounting for both the emissions generated during the burns and the reduction in wildfire smoke when treated areas subsequently reburned. Scaling these efforts to treat one million acres annually, as outlined in California's Wildfire and Forest Resilience Action Plan, could reduce smoke $PM_{2.5}$ emissions by 655,000 metric tons over the next 5 years. However, although we demonstrate that recent Rx fires provide a net benefit by avoiding future wildfire smoke $PM_{2.5}$ emissions, current land management planning in the United States rarely accounts for the averted smoke exposure from wildfires when planning Rx burns on federal lands.

Conflict of Interest

The authors declare no conflicts of interest relevant to this study.

Data Availability Statement

Processed data and Supporting Information S1 used in this work can be found at Zenodo (https://zenodo.org/records/15249372).

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References

Abatzoglou, J. T., Rupp, D. E., O'Neill, L. W., & Sadegh, M. (2021). Compound extremes drive the Western Oregon wildfires of September 2020. *Geophysical Research Letters*, *48*(8), e2021GL092520. https://doi.org/10.1029/2021GL092520

Adlam, C., Almendariz, D., Goode, R. W., Martinez, D. J., & Middleton, B. R. (2022). Keepers of the flame: Supporting the revitalization of indigenous cultural burning. *Society & Natural Resources*, *35*(5), 575–590. https://doi.org/10.1080/08941920.2021.2006385

Afrin, S., & Garcia-Menendez, F. (2021). Potential impacts of prescribed fire smoke on public health and socially vulnerable populations in a Southeastern U.S. state. *Science of the Total Environment*, *794*, 148712. https://doi.org/10.1016/j.scitotenv.2021.148712

Agee, J. K., & Skinner, C. N. (2005). Basic principles of forest fuel reduction treatments. *Forest Ecology and Management*, *211*(1), 83–96. https://doi.org/10.1016/j.foreco.2005.01.034

Anderegg, W. R. L., Chegwidden, O. S., Badgley, G., Trugman, A. T., Cullenward, D., Abatzoglou, J. T., et al. (2022). Future climate risks from stress, insects and fire across US forests. *Ecology Letters*, *25*(6), 1510–1520. https://doi.org/10.1111/ele.14018

Auer, M. R. (2021). Considering equity in wildfire protection. *Sustainability Science*, *16*(6), 2163–2169. https://doi.org/10.1007/s11625-021-01024-8

Beidler, J. L., Baker, K. R., Pouliot, G., & Sacks, J. D. (2024). Encountering prescribed fire: Characterizing the intersection of prescribed fire and wildfire in the CONUS. *ACS EST Air*, *1*(12), 1687–1695. https://doi.org/10.1021/acsestair.4c00228

Boer, M. M., Sadler, R. J., Wittkuhn, R. S., McCaw, L., & Grierson, P. F. (2009). Long-term impacts of prescribed burning on regional extent and incidence of wildfires—Evidence from 50 years of active fire management in SW Australian forests. *Forest Ecology and Management*, *259*(1), 132–142. https://doi.org/10.1016/j.foreco.2009.10.005

Brodie, E. G., Knapp, E. E., Brooks, W. R., Drury, S. A., & Ritchie, M. W. (2024). Forest thinning and prescribed burning treatments reduce wildfire severity and buffer the impacts of severe fire weather. *Fire Ecology*, *20*(1), 17. https://doi.org/10.1186/s42408-023-00241-z

Burke, M., Childs, M. L., de la Cuesta, B., Qiu, M., Li, J., Gould, C. F., et al. (2023). The contribution of wildfire to $PM_{2.5}$ trends in the USA. *Nature*, *622*(7984), 761–766. https://doi.org/10.1038/s41586-023-06522-6

California's Wildfire and Forest Resilience Action Plan. (2021). Recommendations of the Governor's forest management task force. Retrieved from https://wildfiretaskforce.org/wp-content/uploads/2022/04/californiawildfireandforestresilienceactionplan.pdf

Collins, L., Trouvé, R., Baker, P. J., Cirulus, B., Nitschke, C. R., Nolan, R. H., et al. (2023). Fuel reduction burning reduces wildfire severity during extreme fire events in south-eastern Australia. *Journal of Environmental Management*, *343*, 118171. https://doi.org/10.1016/j.jenvman.2023.118171

Coogan, S. C. P., Cai, X., Jain, P., & Flannigan, M. D. (2020). Seasonality and trends in human- and lightning-caused wildfires ≥2 Ha in Canada, 1959–2018. *International Journal of Wildland Fire*, *29*(6), 473–485. https://doi.org/10.1071/WF19129

Crippen, R., Buckley, S., Agram, P., Belz, E., Gurrola, E., Hensley, S., et al. (2016). NASADEM global elevation model: Methods and progress. In *The International Archives of the Photogrammetry, Remote Sensing and Spatial Information Sciences, XLI-B4*. In *XXIII ISPRS Congress, Commission IV (Volume XLI-B4)* (pp. 125–128). https://doi.org/10.5194/isprs-archives-XLI-B4-125-2016

Davis, K. T., Peeler, J., Fargione, J., Haugo, R. D., Metlen, K. L., Robles, M. D., & Woolley, T. (2024). Tamm review: A meta-analysis of thinning, prescribed fire, and wildfire effects on subsequent wildfire severity in conifer dominated forests of the Western US. *Forest Ecology and Management*, *561*, 121885. https://doi.org/10.1016/j.foreco.2024.121885

Deak, A. L., Lucash, M. S., Coughlan, M. R., Weiss, S., & Silva, L. C. R. (2024). Prescribed fire placement matters more than increasing frequency and extent in a simulated Pacific Northwest landscape. *Ecosphere*, *15*(4), e4827. https://doi.org/10.1002/ecs2.4827

Dorph, A., Marshall, E., Parkins, K. A., & Penman, T. D. (2022). Modelling ignition probability for human- and lightning-caused wildfires in Victoria, Australia. *Natural Hazards and Earth System Sciences*, *22*(10), 3487–3499. https://doi.org/10.5194/nhess-22-3487-2022

Eidenshink, J., Schwind, B., Brewer, K., Zhu, Z.-L., Quayle, B., & Howard, S. (2007). A project for monitoring trends in burn severity. *Fire Ecology*, *3*(1), 3–21. https://doi.org/10.4996/fireecology.0301003

El Asmar, R., Li, Z., Tanner, D. J., Hu, Y., O'Neill, S., Huey, L. G., et al. (2024). A multi-site passive approach for studying the emissions and evolution of smoke from prescribed fires. *EGUsphere*, 1–40. https://doi.org/10.5194/egusphere-2024-1485

Fernández-Guisuraga, J. M., & Fernandes, P. M. (2024). Prescribed burning mitigates the severity of subsequent wildfires in Mediterranean shrublands. *Fire Ecology*, *20*(1), 4. https://doi.org/10.1186/s42408-023-00233-z

Fuhlendorf, S., Limb, R., Engle, D., & Miller, R. (2011). Assessment of prescribed fire as a conservation practice. Retrieved from https://www.researchgate.net/profile/Samuel-Fuhlendorf/publication/265000863_Assessment_of_Prescribed_Fire_as_a_Conservation_Practice/links/53fde1290cf22f21c2f85b70/Assessment-of-Prescribed-Fire-as-a-Conservation-Practice.pdf

Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., & Moore, R. (2017). Google Earth engine: Planetary-scale geospatial analysis for everyone. *Remote Sensing of Environment*, *202*, 18–27. https://doi.org/10.1016/j.rse.2017.06.031

H.R.5376 - 117th Congress (2021-2022). (2022). Inflation reduction act of 2022 (2021-09-27) [Legislation]. Retrieved from https://www.congress.gov/bill/117th-congress/house-bill/5376

Hunter, M. E., & Robles, M. D. (2020). Tamm review: The effects of prescribed fire on wildfire regimes and impacts: A framework for comparison. *Forest Ecology and Management*, *475*, 118435. https://doi.org/10.1016/j.foreco.2020.118435

Jin, S., Dewitz, J., Li, C., Sorenson, D., Zhu, Z., Shogib, M. R. I., et al. (2023). National land cover database 2019: A comprehensive strategy for creating the 1986–2019 forest disturbance product. *Journal of Remote Sensing*, *3*, 0021. https://doi.org/10.34133/remotesensing.0021

Jonko, A., Oliveto, J., Beaty, T., Atchley, A., Battaglia, M. A., Dickinson, M. B., et al. (2024). How will future climate change impact prescribed fire across the contiguous United States? *Npj Climate and Atmospheric Science*, *7*(1), 1–10. https://doi.org/10.1038/s41612-024-00649-7

Keiter, R. B. (2012). Wildfire policy, climate change, and the law wildfire law edition. *Texas A&M Journal of Property Law*, *1*(1), 87–108. https://doi.org/10.37419/twjrpl.v1.i1.4

Kelp, M. M., Carroll, M. C., Liu, T., Yantosca, R. M., Hockenberry, H. E., & Mickley, L. J. (2023). Prescribed burns as a tool to mitigate future wildfire smoke exposure: Lessons for states and rural environmental justice communities. *Earth's Future*, *11*(6), e2022EF003468. https://doi.org/10.1029/2022EF003468

Kiely, L., Neyestani, S. E., Binte-Shahid, S., York, R. A., Porter, W. C., & Barsanti, K. C. (2024). California case study of wildfires and prescribed burns: $PM_{2.5}$ emissions, concentrations, and implications for human health. *Environmental Science & Technology*, *58*(12), 5210–5219. https://doi.org/10.1021/acs.est.3c06421

Kolden, C. A. (2019). We're not doing enough prescribed fire in the Western United States to mitigate wildfire risk. *Fire*, *2*(2), 30. Article 2. https://doi.org/10.3390/fire2020030

Kramer, A. L., Liu, J., Li, L., Connolly, R., Barbato, M., & Zhu, Y. (2023). Environmental justice analysis of wildfire-related $PM_{2.5}$ exposure using low-cost sensors in California. *Science of the Total Environment*, *856*, 159218. https://doi.org/10.1016/j.scitotenv.2022.159218

Lake, F. K., & Christianson, A. C. (2019). Indigenous fire stewardship. *Encyclopedia of Wildfires and Wildland-Urban Interface (WUI) Fires*, 1–9. https://doi.org/10.1007/978-3-319-51727-8_225-1

Marsavin, A., Gageldonk, R., Bernays, N., May, N., Jaffe, A., & L. Fry, J. (2023). Optical properties of biomass burning aerosol during the 2021 Oregon fire season: Comparison between wild and prescribed fires. *Environmental Sciences: Atmosphere*, *3*(3), 608–626. https://doi.org/10.1039/D2EA00118G

Miller, J. D., & Thode, A. E. (2007). Quantifying burn severity in a heterogeneous landscape with a relative version of the delta Normalized Burn Ratio (dNBR). *Remote Sensing of Environment*, *109*(1), 66–80. https://doi.org/10.1016/j.rse.2006.12.006

Picotte, J. J., Bhattarai, K., Howard, D., Lecker, J., Epting, J., Quayle, B., et al. (2020). Changes to the monitoring trends in burn severity program mapping production procedures and data products. *Fire Ecology*, *16*(1), 16. https://doi.org/10.1186/s42408-020-00076-y

Radeloff, V. C., Helmers, D. P., Kramer, H. A., Mockrin, M. H., Alexandre, P. M., Bar-Massada, A., et al. (2018). Rapid growth of the US wildland-urban interface raises wildfire risk. *Proceedings of the National Academy of Sciences of the United States of America*, *115*(13), 3314–3319. https://doi.org/10.1073/pnas.1718850115

Rao, K., Williams, A. P., Diffenbaugh, N. S., Yebra, M., Bryant, C., & Konings, A. G. (2023). Dry live fuels increase the likelihood of lightning-caused fires. *Geophysical Research Letters*, *50*(15), e2022GL100975. https://doi.org/10.1029/2022GL100975

Rao, K., Williams, A. P., Diffenbaugh, N. S., Yebra, M., & Konings, A. G. (2022). Plant-water sensitivity regulates wildfire vulnerability. *Nature Ecology & Evolution*, *6*(3), 332–339. https://doi.org/10.1038/s41559-021-01654-2

Rosenberg, A., Hoshiko, S., Buckman, J. R., Yeomans, K. R., Hayashi, T., Kramer, S. J., et al. (2024). Health impacts of future prescribed fire smoke: Considerations from an exposure scenario in California. *Earth's Future*, *12*(2), e2023EF003778. https://doi.org/10.1029/2023EF003778

Schollaert, C. L., Marlier, M. E., & Busch Isaksen, T. M. (2024). Development of a source-specific biomass burning emissions inventory for Washington, Oregon, and California. *Atmospheric Environment*, *319*, 120283. https://doi.org/10.1016/j.atmosenv.2023.120283

Soler, A., Pineda, N., Segundo, H. S., Bech, J., & Montanyà, J. (2021). Characterisation of thunderstorms that caused lightning-ignited wildfires. *International Journal of Wildland Fire*, *30*(12), 954–970. https://doi.org/10.1071/WF21076

Sow, M., Hély, C., Mbow, C., & Sambou, B. (2013). Fuel and fire behavior analysis for early-season prescribed fire planning in Sudanian and Sahelian savannas. *Journal of Arid Environments*, *89*, 84–93. https://doi.org/10.1016/j.jaridenv.2012.09.007

St. Denis, L. A., Short, K. C., McConnell, K., Cook, M. C., Mietkiewicz, N. P., Buckland, M., & Balch, J. K. (2023). All-hazards dataset mined from the US national Incident management System 1999–2020. *Scientific Data*, *10*(1), 112. https://doi.org/10.1038/s41597-023-01955-0

Stephens, S. L., & Moghaddas, J. J. (2005). Experimental fuel treatment impacts on forest structure, potential fire behavior, and predicted tree mortality in a California mixed conifer forest. *Forest Ecology and Management*, *215*(1), 21–36. https://doi.org/10.1016/j.foreco.2005.03.070

Stevens-Rumann, C. S., Kemp, K. B., Higuera, P. E., Harvey, B. J., Rother, M. T., Donato, D. C., et al. (2018). Evidence for declining forest resilience to wildfires under climate change. *Ecology Letters*, *21*(2), 243–252. https://doi.org/10.1111/ele.12889

Swain, D. L., Abatzoglou, J. T., Kolden, C., Shive, K., Kalashnikov, D. A., Singh, D., & Smith, E. (2023). Climate change is narrowing and shifting prescribed fire windows in Western United States. *Communications Earth & Environment*, *4*(1), 1–14. https://doi.org/10.1038/s43247-023-00993-1

United States Government Accountability Office. (2019). Wildland fire: Federal agencies' efforts to reduce wildland fuels and lower risk to communities and ecosystems. Retrieved from https://www.gao.gov/assets/710/703470.pdf

Vaillant, N. M., Fites-Kaufman, J. A., & Stephens, S. L. (2009). Effectiveness of prescribed fire as a fuel treatment in Californian coniferous forests. *International Journal of Wildland Fire*, *18*(2), 165. https://doi.org/10.1071/WF06065

Wen, J., Heft-Neal, S., Baylis, P., Boomhower, J., & Burke, M. (2023). Quantifying fire-specific smoke exposure and health impacts. *Proceedings of the National Academy of Sciences*, *120*(51), e2309325120. https://doi.org/10.1073/pnas.2309325120

Wiedinmyer, C., Kimura, Y., McDonald-Buller, E. C., Emmons, L. K., Buchholz, R. R., Tang, W., et al. (2023). The Fire Inventory from NCAR version 2.5: An updated global fire emissions model for climate and chemistry applications. *Geoscientific Model Development*, *16*(13), 3873–3891. https://doi.org/10.5194/gmd-16-3873-2023

Wu, X., Sverdrup, E., Mastrandrea, M. D., Wara, M. W., & Wager, S. (2023). Low-intensity fires mitigate the risk of high-intensity wildfires in California's forests. *Science Advances*, *9*(45), eadi4123. https://doi.org/10.1126/sciadv.adi4123

Xu, Q., Westerling, A. L., Notohamiprodjo, A., Wiedinmyer, C., Picotte, J. J., Parks, S. A., et al. (2022). Wildfire burn severity and emissions inventory: An example implementation over California. *Environmental Research Letters*, *17*(8), 085008. https://doi.org/10.1088/1748-9326/ac80d0

Zhou, X., Josey, K., Kamareddine, L., Caine, M. C., Liu, T., Mickley, L. J., et al. (2021). Excess of COVID-19 cases and deaths due to fine particulate matter exposure during the 2020 wildfires in the United States. *Science Advances*, *7*(33), eabi8789. https://doi.org/10.1126/sciadv.abi8789